

NOTE:

Cd, Pb AND ORGANOCHLORINE PESTICIDES OF *MYTELLA STRIGATA* (PELECYPODA: MYTILIDAE) OF SIX COASTAL LAGOONS OF NW MEXICO

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RESUMEN

Contenido de Cd, Pb y pesticidas organoclorados de *Mytella strigata* (Pelecypoda: Mytilidae) de seis lagunas costeras del noroeste de México. Los contenidos de Cd y Pb de los tejidos blandos de los mejillones de mangle *Mytella strigata* colectados en 1996, en seis lagunas costeras del noroeste de México variaron entre 0.73 y 1.9 $\mu\text{g g}^{-1}$ y entre 8.3 y 17.1 $\mu\text{g g}^{-1}$, respectivamente; los valores de DDD variaron entre 4.5 y 119 ng g^{-1} , los de DDE desde menos del límite de detección (<DL) hasta 205 ng g^{-1} y el DDT varió entre <DL hasta 4.1 ng g^{-1} , con los valores más altos en las lagunas cercanas a zonas de agricultura intensiva, que se encuentran en la parte central y septentrional del estado de Sinaloa. El intervalo de las concentraciones de HCB fue de <DL hasta 695 ng g^{-1} y los valores mayores se midieron en las lagunas meridionales, en áreas de agricultura extensiva y menos mecanizada. Los valores de Pb superaron los límites aceptables para el consumo humano, mientras que los restantes no alcanzaron niveles preocupantes para la salud humana o ambiental, pero indican una tendencia al aumento en comparación con estudios realizados con anterioridad.

PALABRAS CLAVE: Bivalvos, Pacífico mexicano, Contaminación, Inocuidad alimentaria.

There is growing evidence of the coastal water contamination of the Mexican Pacific and, in spite of the official policies directed to regulate contaminant emissions, this is likely to increase in view of the intense agricultural, industrial and fisheries-related activities of the coastal states of NW Mexico (Soto-Jiménez *et al.*, 2003; González-Farías *et al.*, 2006). These water bodies are nursery and feeding areas for many aquatic species and support several traditional fisheries, as well as important shrimp and mollusc aquaculture developments (De la Lanza-Espino *et al.*, 1994). For this reason, if their contamination was increasing this might be of concern for environmental and human health. However, there is little information on the previous levels of contaminants of these aquatic environments.

Different species of mussels have been used worldwide for several decades as sentinels, to provide information on the levels of contamination of the marine coastal environment (Rainbow, 1995; Cantillo, 1998; Mubiana *et al.*, 2005). This study provides background information on the concentration of some contaminants determined in 1996 in the soft tissues of the mangrove mussel *Mytella strigata* collected in six coastal lagoons of NW Mexico.

The mussels (shell height range: 3 to 4.7 cm) were hand-collected between April and May 1996 from the mangrove roots of two-three sampling sites of the intertidal zone of the lagoons of Navachiste (NA), Altata-Ensenada del Pabellón (AEP), Ceuta (CE), Estero de Urías (EU), Mexcaltitán (MX), and San Cristóbal (SC), located in the Mexican Pacific coastal plains of the states of Sinaloa (NA to EU) and Nayarit (MX and SC: Figure 1). Their surface areas range from 24 ha (SC) to 36,000 ha (AEP), the tidal regime is semidiurnal and the average tidal range of these lagoons is close to 1 m.

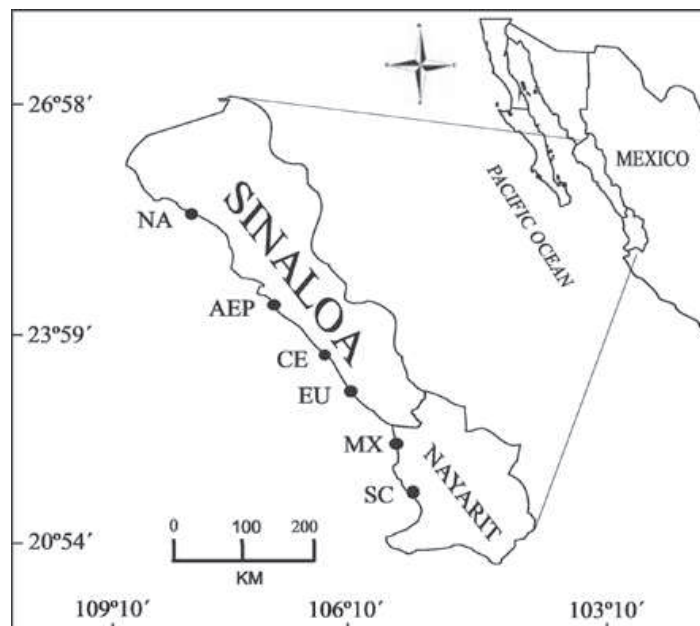


Figure 1. Location of the six lagoons: Navachiste: NA; Altata Ensenada del Pabellón: AEP; Ceuta: CE; Estero de Urías: EU; Mexcaltitán: MX; San Cristóbal: SC.

All these lagoons receive effluents from human activities, mainly of intensive agriculture in the case of NA, AEP and CE. AEP and EU are also the receiving water bodies of the urban and industrial wastewaters of the cities of Culiacán and Mazatlán, and MX and SC are located in areas of less intensive, traditional agriculture and fisheries.

In the laboratory, 30 mussels obtained in each sampling site of each lagoon were washed free of sediments and epibionts and shucked. The pooled samples were freeze-dried, homogenized in a teflon mortar and the respective analyses were performed on three subsamples of these composite samples.

The contaminants determined were two non essential heavy metals (Cd and Pb), and the organochlorine pesticides dichlorodiphenyltrichloroethane (4,4 p,p: DDT), dichlorodiphenyldichloroethane (4,4 p,p: DDD), dichlorodiphenylethylene (4,4 o,p: DDE), and hexachlorobenzene (HCB) determined in a similar study by Páez-Osuna *et al.* (2002). All glassware and materials used for sample handling were acid-cleaned as in Moody and Lindstrom (1977) for metal analysis, or washed with methanol and hexane for pesticides (IDF, 1983). All reagents used for digestion and extraction were metal-free for heavy metals and HPLC-grade for pesticides.

After acid digestion of the subsamples (1 g in 25 mL concentrated HNO₃, evaporation to dryness at 90 °C and solution of the residue in 20 mL 2M HNO₃; Páez-Osuna *et al.*, 2002), trace metals were determined by flame atomic absorption spectrophotometry using the method of internal standard additions, and mussel homogenate MA-M-2/MT as an analytical reference. The respective percentages of recovery of Cd and Pb were 96 and 85 %.

DDT, DDD, DDE and HCB were obtained by solvent extraction from freeze-dried, homogenized tissue samples (10-15 g), separated using a chromatography column of activated florisil and eluted with hexane (fraction 1), hexane/methylene chloride 7:3 v/v (fraction 2), and methylene chloride (fraction 3). Each fraction was concentrated to 1 mL as in UNEP (1988) and the pesticides were determined with a Hewlett Packard 5890 gas chromatograph with electron-capture detectors, using fused-silica capillary columns (25 m x 0.2 mm ID x 0.33 µm, HP). The temperature program was 70 °C for 2 min, increasing by 3 °C min⁻¹ to 260 °C and held constant for 10 min. Quantification and identification of the pesticides were carried out using Supelco™ standards. The efficiency of recovery, determined using 2, 3, 4-tetrachlorobiphenyl as internal standard, ranged from 70 to 100 %.

The highest mean Cd and Pb concentrations were determined in NA and AEP (Cd: 1.9 and 1.7 µg g⁻¹; Pb: 13.0 and 17.1 µg g⁻¹, respectively), whereas the other lagoons had values ranging between 0.81 and 0.73 µg g⁻¹ (Cd) and from 8.3 to 11.0 µg g⁻¹ (Pb). CE and SC shared the lowest values for both metals (Figure 2). Pb

levels are expected to be high in water bodies close to urban or industrial centers, or to other areas of intense human activities and related use of surface vehicles, such as the intensive agricultural and food-related industries of the areas surrounding the NA and AEP lagoonal systems (Green-Ruiz and Páez-Osuna, 2003).

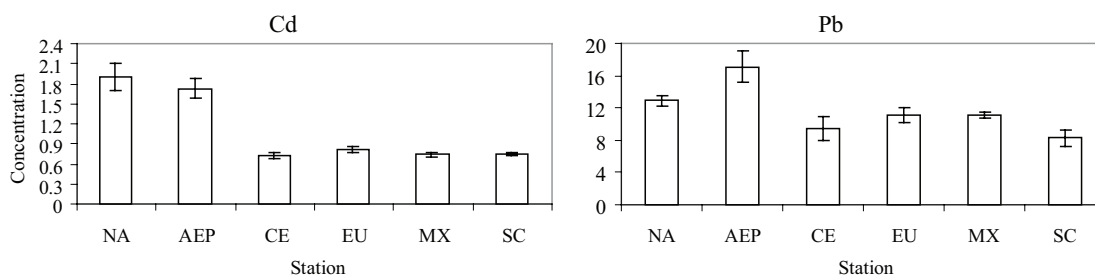


Figure 2. Cd and Pb concentrations in the soft tissues of *Mytella strigata* ($\mu\text{g}\cdot\text{g}^{-1}$, dry weight). Data are mean values \pm sd.

The Cd and Pb concentrations reported in EU (0.8 and $8.1 \mu\text{g}\cdot\text{g}^{-1}$) are almost twice and one order of magnitude higher than the respective values of 0.45 (Cd) and $0.9 \mu\text{g}\cdot\text{g}^{-1}$ (Pb) determined by Szefer *et al.* (1998) for *M. strigata* collected in different areas of the same lagoonal system, although the geographic trend and the range of Pb contents coincide with those found in the same year by Páez-Osuna *et al.* (2002) in the mangrove oyster *C. corteziensis*. According to Cantillo (1998) the values indicative of metal contamination are $3.2 \mu\text{g}\cdot\text{g}^{-1}$ for Cd and $3.7 \mu\text{g}\cdot\text{g}^{-1}$ for Pb. The range of Pb concentrations observed in our specimens was 8.3 - $17.1 \mu\text{g}\cdot\text{g}^{-1}$, all above the threshold value and, considering a water content of between 80 and 85% , all were close or higher than the level considered safe for human consumption ($1.5 \mu\text{g}\cdot\text{g}^{-1}$ wet weight: FDA, 2001, GUCE, 2006) indicating that a decade ago Pb had already reached levels of concern in Mexican coastal water bodies both from the environmental as well as from the food safety points of view.

Bivalves species have specific metal accumulation strategies, and for this reason the use of same species is recommended for comparative purposes. However, there is only one recent study with *Mytella strigata* of the Mexican Pacific coastal waters (Frías-Espericueta *et al.* 2008), who found Cd and Pb values of 6.30 and $5.22 \mu\text{g}\cdot\text{g}^{-1}$ in AEP lagoon.

In addition, Frías-Espericueta *et al.* (2009) reported annual values of 6.47 , 6.91 and $1.55 \mu\text{g}\cdot\text{g}^{-1}$ for Cd, and Pb contents of 4.86 , 8.09 and $6.51 \mu\text{g}\cdot\text{g}^{-1}$ in the mangrove oyster *Crassostrea corteziensis* of the lagoons NA, CE and EU, respectively. According to these data, there seems to be a relative Cd increase and a Pb decrease in the area, probably related in the first case to the intensification of agricultural activities (Berlanga and Ruiz-Luna, 2007) since Cd is associated with

phosphate fertilizers (Loganathan *et al.*, 2008), and in the second because the use of Pb-added gasoline became illegal in Mexico in 1997 (Gavilán-García *et al.*, 2004).

The more than 50 years of intensive agriculture in the northern and central parts of the state of Sinaloa are reflected in the high values of DDT, DDD and DDE determined in AEP, CE and NA, respectively, with 4.1, 119.3 and 205.7 ng g⁻¹; the highest concentration of HCB (695 ng g⁻¹) was that of Estero de Urías, while Mexcaltitán showed the lowest concentrations of all these compounds (Table 1). Until recent years, the production and use of several persistent pesticides were not restricted in Mexico and, even in the case of restrictions as a result of international pressure, these have not been fully implemented and adequately enforced (Albert, 1996).

Table 1. Concentrations of organochlorine pesticides (ng g⁻¹, dry weight). <DL: below detection limit (0.1 ng g⁻¹).

Lagoon	DDT	DDD	DDE	HCB
NA	<DL	76.8	205.7	98.6
AEP	4.1	12.3	83.2	5.2
CE	<DL	119.3	28.3	11.1
EU	<DL	12.0	<DL	695
MX	<DL	4.5	0.2	<DL
SC	2.1	5.3	61.8	<DL

This explains the data of 4.1 ng g⁻¹ of DDT detected in AEP, which is two to three orders of magnitude higher than the 1.8 to 65 ng g⁻¹ and the 5.6 ng g⁻¹ reported by Gutiérrez-Galindo *et al.* (1983, 1988) and Botello *et al.* (1994) for other areas of the NW Pacific coast and the Gulf of Mexico. The same applies to the concentrations of DDD and DDE, which were by far higher than those found in El Salvador by Michel and Zengel (1998). All the high values found in this lagoon may be due to pollutants accumulation generated locally or imported from surrounding areas, caused by its limited water exchange with the open ocean (Carbajal and Núñez-Riboni, 2002).

Only the DDE values found in EU and MX were lower than the 0.075-0.423 µg g⁻¹ of DDE determined by Wade *et al.* (1998) in the coastal zone of SE USA, where HCB concentrations ranged from <0.25-2.4 ng g⁻¹ that, with the exception of SC and MX lagoons, are by far lower than those of this study. Considering that these lagoons are important traditional fishing grounds, that all samples showed high levels of Pb and that the organic pollutants studied are likely to cause at least sublethal effects on a wide range of aquatic organisms at the ng g⁻¹ level (Munn *et al.*, 2006), our results show that more research is needed on the levels of contamination of the

coastal ecosystems of NW Mexico, since the state of pollution of these water bodies of the Mexican Pacific could have already reached levels of serious concern.

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